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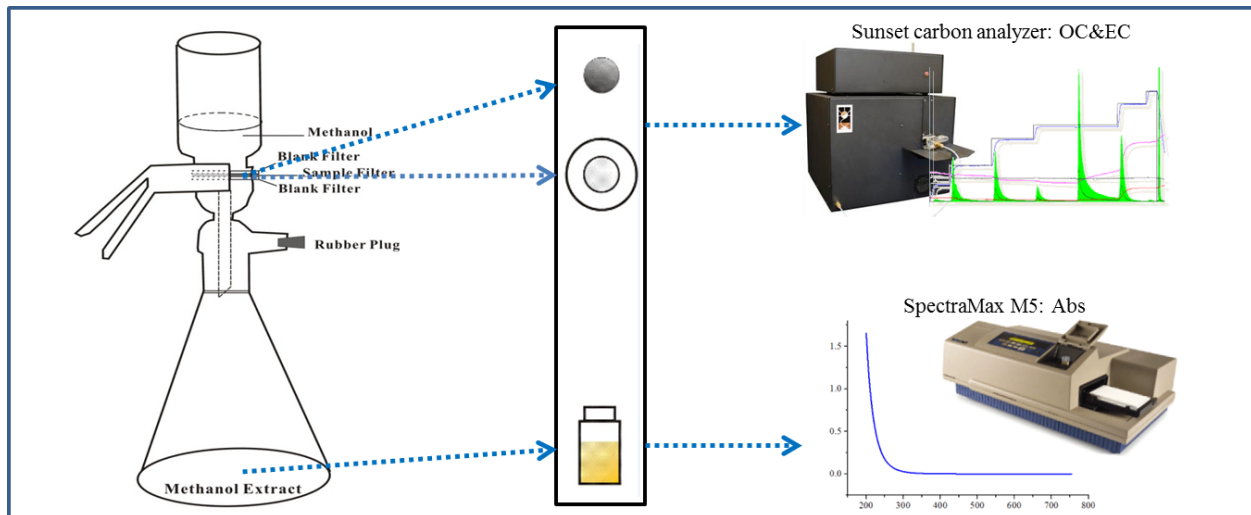
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1 A new method for extraction of methanol-soluble brown carbon:

2 Implications for investigation of its light absorption ability

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18

19 **ABSTRACT**

20 As an important component of organic carbon (OC), brown carbon (BrC) plays a
21 significant role in radiative forcing in the atmosphere. Water-insoluble OC (WIOC)
22 generally has higher light absorption ability than water-soluble OC (WSOC). The
23 mass absorption cross-section (MAC) of WIOC is normally investigated by dissolving
24 OC in methanol. However, all the current methods have shortcomings due to
25 neglecting the methanol insoluble particulate carbon that is detached from the filter
26 and suspended in methanol extracts, which results in MAC uncertainties of the
27 methanol-soluble BrC and its climate warming estimation. In this study, by
28 investigating typical biomass combustion sourced aerosols from the Tibetan Plateau
29 and ambient aerosols from rural and urban areas in China, we evaluated the
30 extractable fraction of OC for the existing methods. Moreover, a new method was
31 developed to overcome the methanol insoluble particulate carbon detachment problem
32 to achieve more reliable MAC values. We found that OC can be dissolved in
33 methanol in a short time (e.g., one hour) and ultrasonic treatment and long-term
34 soaking do not significantly increase the extractable OC fraction. Additionally, we
35 proved that methanol insoluble particulate carbon detachment in methanol does exist
36 in previous methods, causing overestimation of the BrC mass extracted by methanol
37 and thus the underestimation of MAC values. We therefore recommend the newly
38 developed extraction method in this study to be utilized in future related studies to
39 quantitatively obtain the light absorption property of methanol-soluble BrC.

40 **Keywords:** methanol-soluble brown carbon; MAC; biomass aerosol; ambient aerosol

41 **Capsule**

42 A new method is developed to overcome the particulate carbon detachment problem
43 during the extraction of methanol-soluble BrC, and thus the reliable MAC is obtained.

44 **1. Introduction**

45 Aerosol particles are among the largest sources of uncertainties in estimating
46 radiative climate forcing (Anderson et al., 2003). As an important part of aerosol
47 particles, organic carbon (OC) plays an important role in solar radiation absorbing in
48 addition to its solar radiation scattering, especially in the ultraviolet (UV) wavelength
49 range (Laskin et al., 2015; Yan et al., 2018). Recently, increasing attention has been
50 focused on those carbon fractions that absorb solar radiation, i.e., brown carbon (BrC),
51 in both models and in situ observation studies (Andreae and Gelencsér, 2006;
52 Kirillova et al., 2014; Lin et al., 2014; Saleh et al., 2013; Shamjad et al., 2015; Xie et
53 al., 2019). For instance, it is estimated that BrC accounts for approximately 24% of
54 the combined black carbon (BC) and BrC warming effects at the tropopause (Zhang et
55 al., 2017). BrC is generally divided into water-soluble brown carbon (WS-BrC) and
56 water-insoluble brown carbon. Light absorption of the latter carbon fraction is
57 generally evaluated by treating the sample with methanol (Chen and Bond, 2010).
58 Therefore, this BrC fraction is called methanol-soluble brown carbon (MeS-BrC). To
59 date, the method for investigating the light absorption of WS-BrC in aerosol samples
60 has been well established, and little uncertainty exists regarding the value of the mass
61 absorption cross-section of WS-BrC (MAC_{WS-BrC}) because OC dissolved in water can
62 be easily and accurately measured by a TOC analyzer after filtering the extracts (Li et

63 al., 2016b). However, OC dissolved in methanol (i.e., MeS-BrC) cannot be directly
64 measured by this method due to the interference of organic solvent, so that indirect
65 calculations have to be adopted (Chen and Bond, 2010; Cheng et al., 2017; Huang et
66 al., 2018). Moreover, the treatment of samples with methanol is still a complicated
67 issue, and no consensus has been reached. To date, aerosol samples are treated with
68 methanol using multiple methods (e.g., different treatment times, with or without
69 sonication) (method details and comparison are provided in Table 1). However, all
70 existing methods have shortcomings. For example, the methanol insoluble particulate
71 carbon detaching from the filter and entering into the methanol cannot be
72 quantitatively measured, thus the indirect calculation of MeS-BrC in previous studies
73 was conducted by assuming either OC of original filter samples could be completely
74 extracted by methanol or the methanol insoluble particulate carbon detachment is
75 ignorable (Table 1), which caused the overestimation of MeS-BrC and accordingly the
76 underestimation of the mass absorption cross-section of MeS-BrC ($MAC_{MeS-BrC}$)
77 (Chen and Bond, 2010; Cheng et al., 2016). In addition, sonication treatment
78 generally increases the extractable fraction of carbon compared to other methods
79 without sonication treatment (Polidori et al., 2008); thus, both $MAC_{MeS-BrC}$ and
80 MAC_{WS-BrC} values of the samples treated by sonication should be higher than those
81 without sonication. Currently, a comparison of methanol extractable OC fraction by
82 the three existing methods presented in Table 1 is still lacking. Because mass
83 absorption cross-section of BrC (MAC_{BrC}) data are the basic input of radiative models
84 of aerosol particles, the aforementioned limitations and uncertainties will ultimately

85 be propagated to climate forcing estimation. Therefore, it is urgent to establish a
86 reliable method for MAC_{BrC} to better evaluate the radiative forcing of carbonaceous
87 particles.

88 Normally, the BrC derived from biomass combustion sourced aerosols is more
89 easily dissolved in water and methanol than that derived from fossil-sourced and
90 ambient aerosols (Chen and Bond, 2010). In this study, $PM_{2.5}$ samples (particulate
91 matter with a diameter smaller than $2.5 \mu m$) originated from yak dung combustion,
92 from a typical urban city on the Tibetan Plateau and from rural areas in East China
93 (which is seriously polluted) were collected (Table S1). The collected samples were
94 treated with methanol by the previous three methods to investigate the difference in
95 the methanol extractable OC fractions among these methods. On this basis, a new
96 method is developed to provide more reliable $MAC_{MeS-BrC}$ values for aerosol samples.

97 Table 1 Information and comparison of three previous extraction methods and the one in this study.

Method	Extraction process	Calculation of MeS-BrC	Potential problems	Reference
Method 1	Filter samples were immersed in methanol for 1 h without shaking or sonication	MeS-BrC=OC of original filter–OC remaining on methanol extracted-filter	The methanol insoluble particulate carbon detachment was ignored	(Cheng et al., 2017)
Method 2	Filter samples were sonicated for 1 h in methanol	MeS-BrC=OC of original filter	The OC of original filter samples was assumed to be completely extracted by methanol	(Huang et al., 2018)
Method 3	Filter samples were sonicated in methanol for 1 h, the solution was kept at room temperature for 20 h, then it was again sonicated for another 1 h	MeS-BrC=OC of original filter–OC remaining on methanol extracted-filter	The methanol insoluble particulate carbon detachment was ignored	(Chen and Bond, 2010)
New method	Filter samples were placed in the middle of a sandwich filtration assembly as presented in Figure 1, and extracted with methanol by three filters is 1 h	MeS-BrC=OC of original filter–OC remaining on middle and bottom methanol extracted filters in sandwich filtration assembly, i.e., equation (2)		This study

99 **2. Methodology**

100 **2.1. Field sites**

101 $PM_{2.5}$ filter samples were collected at four sites in China, namely, Nam Co,
102 Lhasa, Zhangbo and Yangdian (Table S1). Nam Co is a typical remote area on the
103 Tibetan Plateau where local residents burn yak dung for heating and cooking. The
104 aerosol particles emitted from these combustion activities have an important
105 contribution to glacier retreat on the Tibetan Plateau (Li et al., 2016a). Several studies
106 have been conducted on the OC and element carbon (EC) ratios and MAC_{WS-BTC}
107 values at this site (Chen et al., 2015; Hu et al., 2017). Lhasa is the largest city on the
108 south-central Tibetan Plateau; the aerosols in this city are greatly influenced by fossil
109 combustion (Li et al., 2016a). Zhangbo and Yangdian are two typical rural areas in
110 Guanzhong and Huabei Plain, respectively, with serious air pollution in East China.

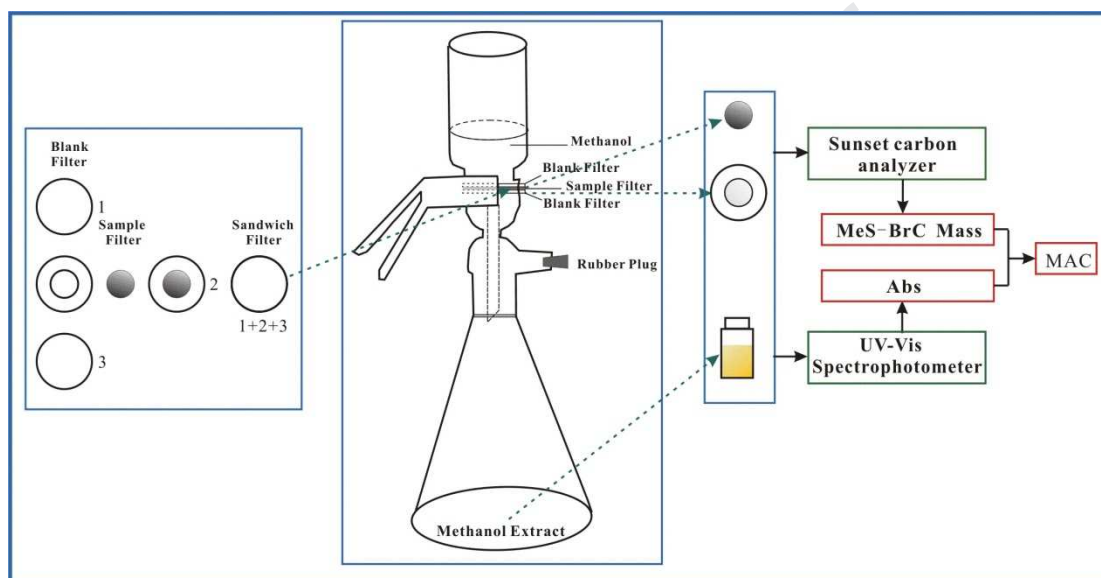
111 The $PM_{2.5}$ samples collected at Nam Co were directly emitted from the yak dung
112 combustion plumes (Chen et al., 2015), which were classified as biomass aerosols.
113 The $PM_{2.5}$ samples collected in the ambient environment at other three sampling sites
114 were classified as ambient aerosols. Due to different sources, the light absorption
115 characteristics of aerosol in two categories are different.

116 **2.2. Extract assembly**

117 A total of 33 aerosol samples were collected for the experiments (Table S1). The
118 filter samples from the urban and rural sites were acidified by fumigation with 37%
119 HCl for 24 h to remove carbonate carbon which interferes the measurement of EC and
120 OC by the carbon analyzer before the experiments followed the practical protocols

121 provided in previous studies (Bosch et al., 2014; Chen et al., 2013; Li et al., 2016a;
122 Pio et al., 2007). The carbonate carbon removed by HCl fumigation accounted for
123 approximately $5.5\pm 2.6\%$ of the total OC of the study aerosols. The analytical
124 procedure in this study is presented in Figure S1. First, a subsample (punch #1) of the
125 original filter was cut out for the measurement of initial OC mass. Second, a 3.8 cm^2
126 subsample (punch #2) was cut out and treated with 38 mL methanol using a sandwich
127 filter assembly to prevent the loss of particles from the filter (Figure 1). In summary,
128 the subsample was punched and placed in a pre-combusted quartz filter (pore size:
129 $0.45\text{ }\mu\text{m}$, Pall TissuquartzTM) with an opening of the same area. The reassembled
130 filter samples were then placed between two pre-combusted quartz filters with the
131 bottom of the subsample facing upward. The thusly formed three-filter sandwich was
132 placed on the glass sand core funnel of a vacuum filtration assembly as presented in
133 Figure 1. After blocking the funnel, 38 mL methanol was added to the filter in three
134 times to keep the long residence time, and the total residence time of the methanol
135 lasted for approximately 1 h to ensure a high OC fraction extracted by methanol. The
136 remaining methanol was then pumped through the filter. Thereafter, the filters were
137 dried in glass petri dishes at $60\text{ }^\circ\text{C}$ for two hours. Punches of 0.526 cm^2 were obtained
138 from the center of the original filter samples, the methanol-treated filter subsamples
139 and the bottom filters of the sandwich filtration assembly. OC and EC of the punched
140 samples were measured using a thermal-optical transmittance (TOT) carbon analyzer
141 (Sunset Laboratory, Tigard, OR, USA) following the IMPROVE-A protocol. The sum
142 of the OC masses of the two punches from the sandwich filtration assembly was

143 regarded as the methanol-insoluble OC fraction. The difference between the OC mass
 144 of the original sample and the methanol-insoluble OC is the mass of the OC dissolved
 145 in methanol, which was used in the calculation of $MAC_{MeS-BrC}$. Third, a subsample
 146 (punch #3) was sonicated in ultrapure water to obtain the MAC_{WS-BrC} value. The
 147 detailed method has been described in our previous article (Li et al., 2016c).
 148



149
 150 **Figure 1.** Filtration unit developed in this study to extract MeS-BrC.

151 2.3. Light absorption measurement and calculation

152 The light absorption spectra of WS-BrC and MeS-BrC were measured using an
 153 ultraviolet-visible (UV-Vis) absorption spectrophotometer (SpectraMax M5, USA), by
 154 scanning from 200-700 nm with a step size of 5 nm. The MAC value was calculated
 155 based on the Beer-Lambert law (Bosch et al., 2014; Kirillova et al., 2014):

$$156 \quad MAC = \frac{Abs}{C \cdot L} \times \ln(10) \quad (1)$$

157 where Abs is the light absorbance measured directly by the spectrophotometer, L is
 158 the absorbing path length (1 cm); C is the mass of OC dissolved in methanol

159 (MeS-BrC) or ultrapure water (WS-BrC).

160 The mass of WS-BrC was directly measured by the TOC analyzer, while the
 161 mass of MeS-BrC was obtained by indirect calculation due to the interference of
 162 methanol as follow:

$$163 \quad \text{MeS-BrC} = \text{OC}_{\text{original}} - \text{OC}_{2,\text{sandwich}} - \text{OC}_{3,\text{sandwich}} \quad (2)$$

164 where $\text{OC}_{\text{original}}$ is the total OC mass of original filter; $\text{OC}_{2,\text{sandwich}}$ and $\text{OC}_{3,\text{sandwich}}$ are
 165 the OC mass of middle and bottom filters extracted by methanol in the sandwich
 166 filtration assembly present in Figure 1.

167 Under the assumption that the OC soluble in water can also be dissolved in
 168 methanol, MAC of the water-insoluble OC (MAC_{WIOC}) can be calculated as the
 169 following ratio:

$$170 \quad \text{MAC}_{\text{WIOC}} = \frac{[\text{Abs}_{365,\text{methanol}} - \text{Abs}_{365,\text{water}}]}{([\text{MeS-BrC}] - [\text{WS-BrC}]) \cdot L} \times \ln 10 \quad (3)$$

171 where $\text{Abs}_{365,\text{methanol}}$ and $\text{Abs}_{365,\text{water}}$ are the light absorbance measured at 365 nm of
 172 filtered methanol and water extracts; $[\text{MeS-BrC}]$ and $[\text{WS-BrC}]$ are the mass of
 173 MeS-BrC and WS-BrC.

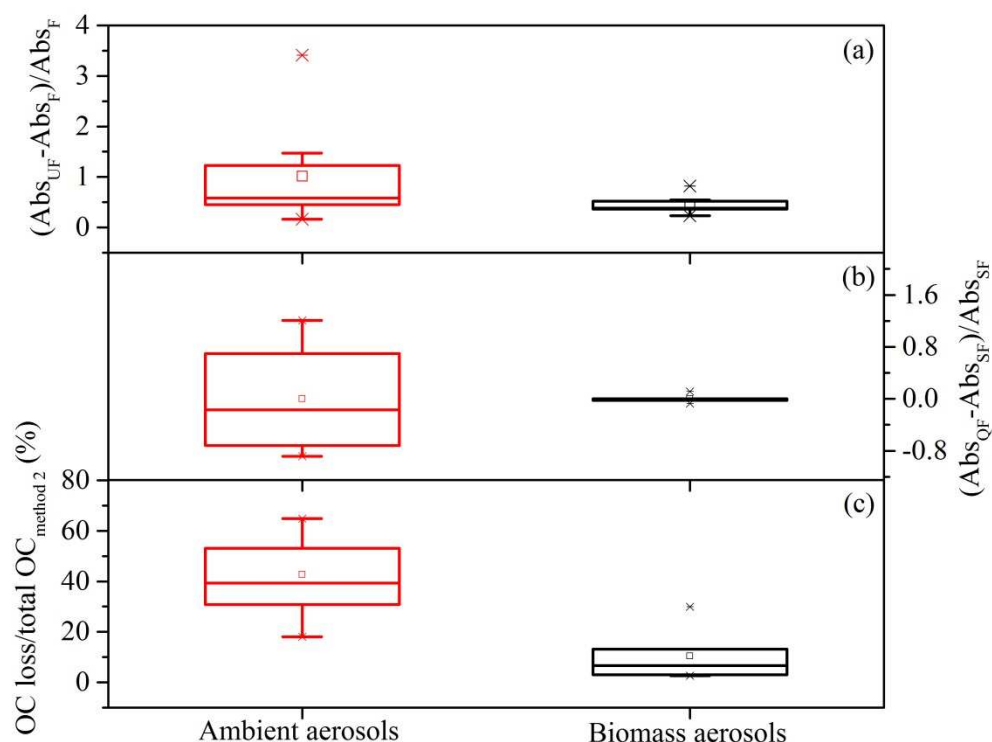
174 **3. Results and discussion**

175 **3.1. The loss of methanol insoluble particulate carbon in three previous methods**

176 The loss of methanol insoluble particulate carbon from the filter during the
 177 methanol extraction was considered insignificant and negligible in a previous study
 178 (Cheng et al., 2017), although it has been reported that this phenomenon will cause
 179 $\text{MAC}_{\text{MeS-BrC}}$ underestimation (Chen and Bond, 2010; Kirillova et al., 2016). Herein, to
 180 evaluate the loss of methanol insoluble particulate carbon into methanol from the

181 filter, light absorption of the methanol extracts using the third method in Table 1 was
182 measured before and after filtration. If the light absorbance at wavelength of 365 nm
183 (Abs_{365}) of the methanol extract after filtration was lower than that before filtration,
184 the loss of methanol insoluble particulate carbon from the filter must occur. Although
185 this approach provides an indirect estimate, the particle loss can thus be evaluated
186 effectively. For instance, due to the detachment of methanol insoluble particulate
187 carbon, Abs_{365} for the unfiltered extracts of ambient and biomass aerosols increases
188 approximately 101% and 45%, respectively compared to those of filtered extracts
189 (Figure 2a), suggesting that the loss of the methanol insoluble particulate carbon does
190 exist in previous methods. Moreover, by collecting these detached OC and those
191 remaining OC on methanol-extracted filters together and measuring by TOT carbon
192 analyzer, the mass of MeS-BrC could be overestimated by 10% and 43% for the
193 biomass and ambient extracts, respectively (Figure 2c) if we assume methanol
194 completely extracts OC of the filter samples (as presented by Huang et al., 2018, i.e.,
195 method 2 in Table 1). In contrast, the new extraction method developed in this study is
196 reliable for obtaining accurate MeS-BrC mass and thus reliable $MAC_{MeS-BrC}$ values by
197 preventing the detachment of the methanol insoluble particulate carbon, which is also
198 verified by the similar average Abs_{365} values compared with the previous filtration
199 methods (Figure 2b). Additionally, the larger Abs increase of the ambient aerosol
200 extracts due to the detachment of methanol insoluble particulate carbon might be
201 explained by the lower fraction of MeS-BrC (i.e., a low ratio of methanol-soluble OC
202 to total OC of the original filter sample) and the existence of large mineral particles

203 with an easy detachment inclination which can be the coating of carbonaceous matter
 204 in ambient aerosols (Figure S2); meanwhile, the finer biomass combustion carbon
 205 particles adhere to the filter paper more easily compared to the ambient particles.
 206 Thus, failing to consider the particle loss would cause large uncertainties in MAC
 207 estimation, especially for the ambient aerosols.

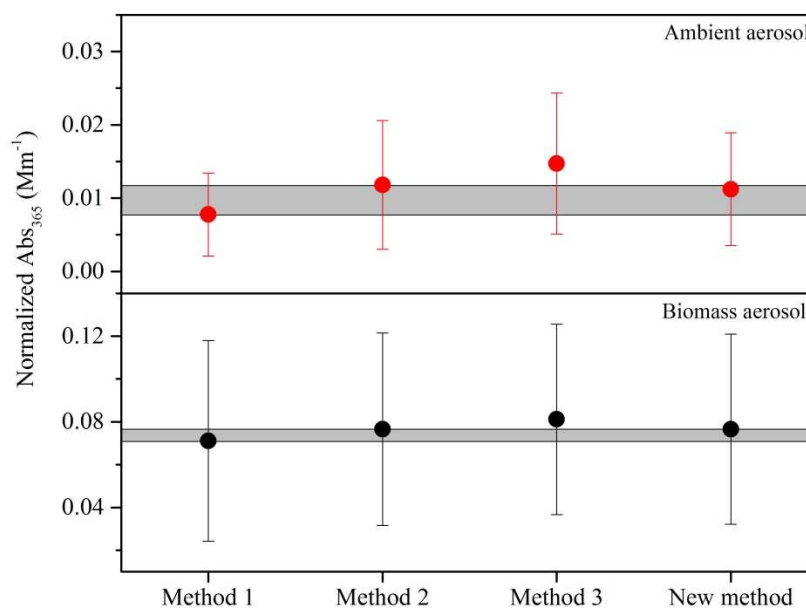


208
 209 **Figure 2.** The increase ratios of Abs_{365} of the unfiltered (Abs_{UF}) methanol extracts
 210 compared to those of Abs_{365} of the filtered extracts (Abs_F) (a), the increase ratios of
 211 Abs_{365} of the extracts filtered by a quartz filter (Abs_{QF}) compared to those of Abs_{365} of
 212 the extracts filtered a by syringe filter (Abs_{SF}) (b), and the ratio of OC loss to total OC
 213 of original filter samples using method 2 in Table 1.

214 3.2. Abs values of MeS-BrC by four different methods

215 The MeS-BrC in the ambient and biomass aerosols was extracted using the
 216 existing three methods (Table 1 and Text S1) and the method we developed in this
 217 study. The Abs_{365} values were measured and compared. The results indicated that the

218 Abs₃₆₅ values of MeS-BrC using the new method were between those obtained using
219 the first two methods but slightly lower than those obtained using the third method
220 (Figure 3). Moreover, the differences in Abs₃₆₅ among the methanol extracts from
221 these four methods were statistically insignificant for both ambient and biomass
222 aerosols ($p>0.05$, $N=33$) (Table S2). Therefore, we propose that the complex
223 treatments of sonication and a longer residence time do not significantly contribute to
224 a larger amount of carbon extracted. Additionally, it is demonstrated that the syringe
225 filter in the previous three methods and the quartz filter paper in this new method do
226 not cause significant differences in the Abs₃₆₅ values (Figure S3). Consequently, it is
227 reasonable to adopt the new method developed in this study to investigate the light
228 absorption of MeS-BrC quantitatively and accurately. However, it should also be
229 pointed out that the theoretical mass absorption cross-section at 365 nm (MAC₃₆₅) of
230 the methanol extracts based on the Abs₃₆₅ values obtained using the third method
231 should be the highest among these four methods due to the elevated Abs₃₆₅ values as a
232 result of the sonication treatment, although the increase is statistically insignificant
233 ($p>0.05$, $N=33$). In addition, the theoretical MAC₃₆₅ divergence of the ambient
234 aerosols between the third method and the new method would be larger than that of
235 the biomass aerosols, which may also indicate that the OC in biomass aerosols is
236 more easily extracted by methanol even without sonication treatment, but sonication
237 treatment has a relatively large effect on OC fraction extracted by methanol in ambient
238 aerosols.



239

240 **Figure 3.** Normalized light absorbance of MeS-BrC at 365 nm by the four different
 241 extraction methods. Note: the large standard deviation (SD) for mean values was attributed to
 242 the large variation in the particulate carbon mass in each collected sample.

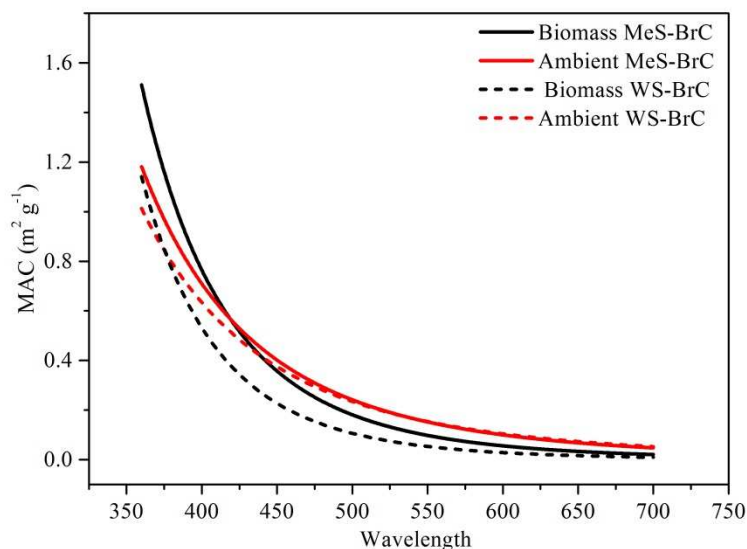
243 3.3. WS-BrC versus MeS-BrC and the light absorption properties

244 Approximately $93.0 \pm 3.8\%$ (87-96%) of the biomass combustion sourced OC
 245 could be extracted within 1 h by methanol in this study, which is likely the lower
 246 bound of the previously reported values of 92%-98% for wood burning aerosols
 247 (Chen and Bond, 2010). Meanwhile, the average methanol extractable OC fraction of
 248 the ambient aerosols was approximately $79.3 \pm 10.0\%$ (55%-93%) (Figure S2), which
 249 was lower than the average value of 89% for the ambient aerosols in Beijing (Cheng
 250 et al., 2017). This phenomenon was possibly because, on the one hand, the aerosol
 251 sources were different between this study and the previous ones, and on the other
 252 hand, the previous two studies failed to consider the methanol insoluble particulate
 253 carbon detachment from the filter during analysis. Correspondingly, the extractable

254 OC fractions by ultrapure water were small ($63.3\pm 13.3\%$ and $55.2\pm 29.1\%$ for the
255 biomass and ambient aerosols, respectively), close to the previously reported values
256 (Cong et al., 2015; Graham et al., 2002; Srinivas and Sarin, 2013). All these results
257 disprove the assumption in several previous studies that methanol extracts almost all
258 the OC from filters (Huang et al., 2018; Kirillova et al., 2016; Zhu et al., 2018), thus
259 causing $MAC_{MeS-BrC}$ underestimation, especially for the ambient aerosols.

260 By using the new method, the MAC_{365} values of the methanol-extracted biomass
261 OC and ambient OC were $1.55\pm 0.43\text{ m}^2\text{ g}^{-1}$ ($1.07\text{-}2.44\text{ m}^2\text{ g}^{-1}$) and $1.14\pm 0.50\text{ m}^2\text{ g}^{-1}$
262 ($0.55\text{-}1.51\text{ m}^2\text{ g}^{-1}$), respectively, in this study. The MeS-BrC fraction has higher MAC
263 values than the WS-BrC fraction across the near-UV wavelength range (Figure 4),
264 consistent with previous results (Chen and Bond, 2010; Kirchstetter et al., 2004; Sun
265 et al., 2007; Zhang et al., 2013). It is clear that the MAC_{365} value of OC from biomass
266 combustion sourced aerosols were higher than that of OC from ambient aerosols;
267 moreover, the MAC_{365} values in Lhasa were lower than those at the other two sites
268 (Table S3) because the fossil fuel contribution to the aerosols in Lhasa was larger than
269 those of the other two sites (Li et al., 2016b). Furthermore, the OC/EC ratios in Lhasa,
270 Zhangbo and Yangdian were 3.4, 7.1 and 7.3, respectively, reflecting the larger fossil
271 fuel contribution of the Lhasa aerosol as reported in previous studies (Hu et al., 2017;
272 Li et al., 2016b). Therefore, it is crucial to obtain the reliable light absorption of OC in
273 the atmosphere not only in the heavily polluted regions but also in the background
274 remote regions of the Tibetan Plateau because there contains the largest number of
275 glaciers outside the polar regions which has been experiencing dramatic retreat due to

276 the climate change and the deposition of light-absorbing impurities (Kang et al., 2019).
 277 Thus, quantitative investigation of the light-absorbing OC and its effect on glaciers
 278 should be taken seriously.

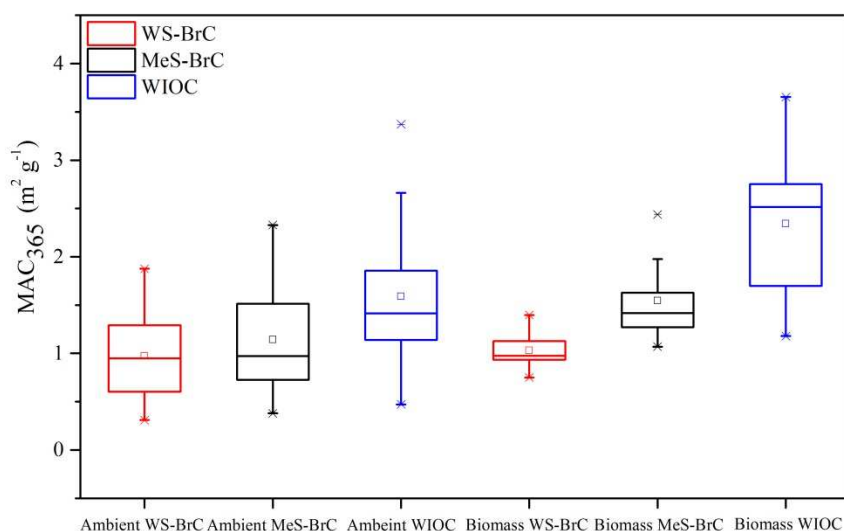


279 **Figure 4.** Average MAC spectra of WS-BrC and MeS-BrC in the ambient and
 280 biomass aerosols.
 281

282 The OC compounds that were only extractable by methanol (i.e., WIOC) were
 283 more light-absorbing than the WS-BrC compounds (Figure 5 and Table S3). The
 284 estimated average MAC_{WIOC} values at 365 nm for the biomass and ambient aerosols
 285 were $2.34 \pm 0.78 \text{ m}^2 \text{ g}^{-1}$ and $1.59 \pm 0.76 \text{ m}^2 \text{ g}^{-1}$, approximately 2.3 and 1.6 times of the
 286 estimated MAC_{WS-BrC} values, respectively. Previous studies have suggested that this
 287 strong light-absorbing WIOC are likely polycyclic aromatic hydrocarbons with a large
 288 molecule weight or certain large molecules containing conjugated aromatic rings
 289 (Apicella et al., 2004; Chen and Bond, 2010; Zhang et al., 2013), which could be
 290 produced during fossil fuel and biomass combustion in both flaming and smoldering
 291 combustion states (Evans and Milne, 1987; Schauer et al., 2001). Correspondingly,

292 the less light-absorbing WS-BrC compounds were identified as humic-like substance
 293 and protein-like substances (Sun et al., 2007; Wu et al., 2019).

294 As a BrC proxy, Ab_{365} of WIOC also exhibited a strong correlation with WIOC
 295 mass in both the ambient and biomass aerosols besides the significant correlation
 296 between WS-BrC and its Abs value (Figure S4), which suggested that a significant
 297 fraction of WIOC is consisted of BrC chromophores. Therefore, the study of
 298 MeS-BrC in aerosols can facilitate a better understanding of its role in climate forcing.
 299 In addition, the new method in this study provides a reliable and quantitative
 300 measurement for this kind of MeS-BrC.



301

302 **Figure 5.** Average MAC_{365} of the water and methanol extracts of ambient and the
 303 biomass aerosols.

304 4. Conclusion

305 By comparing the sandwich filtration assembly developed in this study with the
 306 three previous methods, we proved that the three existing methods do experience the

307 detachment of methanol insoluble particulate carbon from the filter during extraction,
308 which causes large overestimation in the calculation of the MeS-BrC mass and
309 eventually leads to the underestimation of $MAC_{MeS-BrC}$. Correspondingly, the
310 sandwich filtration method provides not only comparable light absorption data of the
311 methanol extracts with the existing methods using traditional sonication process but
312 also the reliable amount of the methanol-extractable OC. Thus, the sandwich filtration
313 method can be utilized to measure the light absorption ability of the methanol-soluble
314 OC in atmospheric aerosols more accurately by overcoming the particulate carbon
315 loss problem.

316 **Note**

317 The authors declare no competing financial interest.

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Highlights

- Light absorption of water-insoluble BrC is studied by extracting OC with methanol
- Uncertainties of previous methods to extract methanol-soluble BrC are discussed
- A new extraction method is developed to obtain reliable MAC of methanol-soluble BrC

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